

# Investigation of Leachate Migration from Dumpsite and Its Effect on Surrounding Groundwater: A Case Study of Kuje Dumpsite, Fct, Abuja, Nigeria

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## ABSTRACT

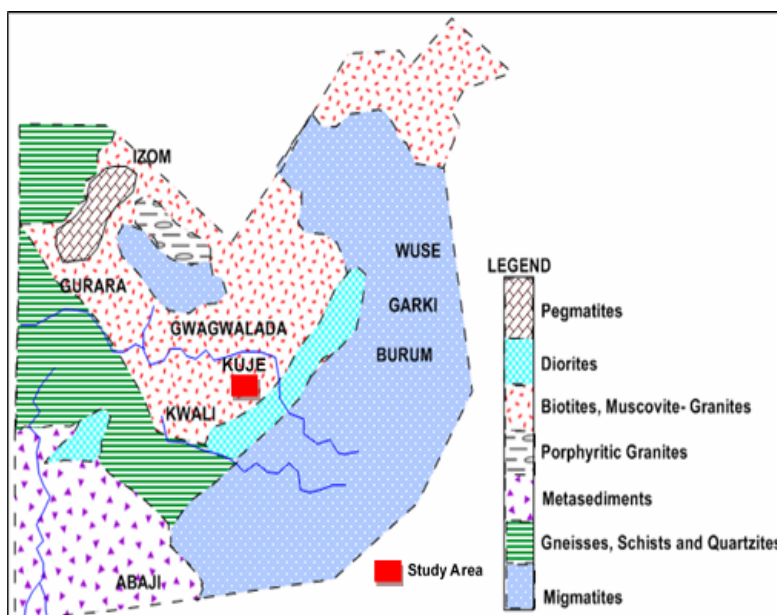
Groundwater contamination is one of the major problems caused by open dumpsite in an environment which directly results into health hazards for the people living around the area. In this study, geophysical and physicochemical methods were carried out to investigate groundwater pollution at a dumpsite in Kuje, FCT, Abuja. From the result of geophysical method, high conductivity of  $0.143 \Omega\text{-1m}$  and low resistivity of  $<7 \Omega\text{m}$ , shows the presence and migration of leachate from the dumpsite. Also, physico-chemical analysis of hand dug well and borehole

water samples were carried out for heavy metal components such as zinc, copper, manganese and iron. The concentrations of heavy metal were high with values of 0.1352, 1.7, 0.36 and 0.32530 respectively for hand dug well, 0.28, 1.65, 0.28 and 0.1387 respectively for borehole water sample, which is above WHO safe limits of 5, 0.05, 0.2 and 0.03 respectively (except for zinc value which is lower than the WHO safe limit). The geophysical and physicochemical results indicate that the refuse at the dumpsite consists of empty cans and rusted iron which cause groundwater pollution. Hence there is a need for proper regulation of the dumpsite to improve water quality, environmental and health standards of the inhabitants of this community.

**Keywords:** Contamination, Dumpsite, Geophysical, Groundwater, Investigation, Physico-Chemical

The study area, Kuje, is located in Abuja's Federal Capital Territory's southern region. The study region is located between longitude  $07^{\circ} 14' 24''\text{E}$  and  $07^{\circ} 14' 41''\text{E}$  and latitude  $08^{\circ} 53' 24''\text{N}$  and  $08^{\circ} 53' 47''\text{N}$ .

## I. INTRODUCTION



**Geological map of FCT showing the study area  
(modified after Obaje, 2009)**

Wastes are items that are undesired. Any item that is thrown away after its initial use or that is useless, broken, or otherwise unusable is considered waste. Solid waste generation is a regular result of necessary human activities. Despite the level of global industrialization and technological advancement that exists today, the intensity of human activity has led to an increase in solid waste globally. According to Abu and Al-Kofahi (2001), landfills have long been used as the final resting place for all waste types. The term "landfill" refers to a facility for the engineered disposal of garbage that uses specialized pollution control techniques to reduce any potential effects. Typically, open landfills are either above ground or located within of pits or quarries.

One of the most important problems with landfill design, operation, and long-term maintenance is leachate management, which happens when water percolates through the waste put in a landfill. Landfills are sources of groundwater and soil pollution because of the leachate that trash migration produces (Misra and Mani, 1991). The simultaneous interaction of physical, chemical, and biological processes results in the overall breakdown of the wastes. Chemically laden leachates, which are one of the results of all these mechanisms and accumulate at the bottom of the dump, are made up of a variety of organic and inorganic compounds. Liners are installed at the base of contemporary landfills to prevent leachate

from migrating there. The migration of leachate into an aquifer is, however, acknowledged to be prevented by such liners, although with time, these liners degrade and finally fail to do so (Kirlna, 2002). Leachate generation at a landfill is influenced by a number of variables, including the climate, site topography, potential landfill cover material, vegetation cover, and type of garbage in the dump. These variables may also indirectly affect the risk of severe water resource impacts. The climate at the site significantly influences the rate of leachate generation, all other factors being equal. Areas near dumpsites have a greater possibility of groundwater contamination because of the potential pollution source of leachate originating from the nearby site (Lokmanet al, 2014). The degree of threat is strongly influenced by the composition of the wastes in the landfill and the volume of leachates generated, as well as the location of the landfill from water bodies; groundwater and surface water (Chrestensen, 2001).

These leachates are a major cause of aquifer contamination because they result from the dumping of home and commercial solid wastes. In studying the leachate effects, it is expected that in areas of high concentrations of leachate, there will be a corresponding very high value of conductivity and a very low value of resistivity (Onyekuru et al, 2018). Since resistivity is the inverse of conductivity, electrical resistivity method is a popular tool for ground water exploration, it is also

used in determining groundwater quality, i.e. whether the water is saline, fresh or contaminated.

Annual waste generation is increasing exponentially with rapid population growth, urbanization and industrial development. The dumping of non-segregated solid waste to landfill sites is the most prevalent waste disposal practice in Nigeria and even in some developed countries (Ferronato et al., 2019). The improper management of landfills and generation of toxic leachate thereby exerts significant impacts on surrounding freshwater and groundwater (Ferronato et al., 2019).

Leachate is the aqueous effluent generated from solid waste owing to their physical, chemical, and biological alteration in landfills (Zhao 2018) and is considered as a chemical soup of dissolved organic matter (DOM), xenobiotic organic compounds, different anions and cations, and heavy metals (Christensen 2001). Among the several elements that make up landfill leachate, heavy metals are non-biodegradable, capable of lowering the quality of surface and groundwater, and hazardous to biological systems even at low concentrations. Heavy metals are also persistent, bio-accumulative, and toxic (Mahmud et al., 2022). Conversely, the DOM which constitute a large portion of leachate, has potential to bind with heavy metal, and consequently plays a significant role in the bio-availability of those metals in the aquatic environments (Baun and Christensen 2004; Parvin et al., 2021). The principal concern about municipal landfill is focused on the pollution potential due to mobilization of the generated leachate through the subsoil into the surface and groundwater. Further, during the wet season, water containing leachate from landfill site drains into the nearby lowlands and surface water bodies and pollutes the local environments. Hence, this toxic aqueous effluent from landfill site can cause potential risk to surface and groundwater (Christensen 2001) and are eventually found to pose a threat for aquatic biota, plant and public health. Iswa, (2013) reported that in most of the developing countries, improperly managed open landfill sites are more commonly practiced than controlled and engineered landfills. Residents, especially the urban and semi-urban residents in those countries are affected severely by this uncontrolled management

of waste via water and food contamination by toxic leachate.

Kuje Area Council is dominated by farmers who specialize in agriculture and livestock breeding based on the information gathered within the community. However, the economic activities of the people are basically fresh meats and other farm products. Thus, solid, liquid and gaseous wastes are dumped or discharged into the dumpsite which could affect soil and groundwater. The sighting of borehole as the only source of good water in this area has become a serious challenge. This challenge is worsened by the fact that there are inadequately trained waste disposal personnel and equipment, poor waste collection, sorting and disposal without regards to the local geology and hydrogeology of the study area, as a result of the imminent impact of solid waste on the environment, it is necessary to investigate the potential for the contamination of groundwater around the dumpsites. The aim of this study is to investigate leachate migration from dumpsite and its effect on surrounding groundwater in the study area to determine depth and thickness of the geoelectric layers, to detect and map contaminated zones, to determine the level and extent of contamination to the aquifer by leachate from the dumpsites and to compare the difference in the quality of sampled water with WHO and NSDWQ water quality standards.

## II. MATERIALS

The materials used for the research include Tarrameter (SAS 300), Global Positioning System (GPS), and Measuring tape, Hammers, Connecting Cables, Electrodes, Extra-large Umbrella, RES2DINV and IPI2win Software

## III. METHOD

The vertical electric sounding (VES) method of electrical resistivity was employed using the Schlumberger electrode configuration. The current electrode spacing AB/2 varied from 1.5 to 150 m was used with the aim of probing a depth of at least 1/3 of AB. Tarrameter (SAS 300) was used in the data acquisition. In this configuration, the electrodes are positioned symmetrically along a straight line, the current electrodes on the outside and the potential electrodes on the inside. To change the depth range of the measurements, the current electrodes were displaced outwards while the potential electrodes in general, were left at the same position. When the ratio of the distance between the current electrodes to that between the

potential electrodes becomes too large, the potential electrodes must also be displaced outward otherwise the potential difference becomes too small to be measured with sufficient accuracy. The current electrodes been shifted to new position for most readings while potential electrodes are kept constant for up to three or four readings, during the field work, taking a sounding, the Tarrameter(SAS 300) performs automatic recording of both voltage and current, stacks the results, computes the resistance in real time and digitally displays it. Other instruments used includes; metal electrodes, measuring tape, hammer (used in hammering the electrodes into the ground), and connecting cables. The field data obtained was subjected to analysis and interpretation by computer iterations using RES2DINV and IPI2win software. The values obtained were first manually plotted against their respective current-electrode spacing values on a log-log graph, partially curve-matched and then electronically iterated using RES2DINV and IPI2win computer iteration program to obtain the geo-electric layers which show vertical variation in resistivity values with depth.



Study Area showing the Profile lines.

#### IV.

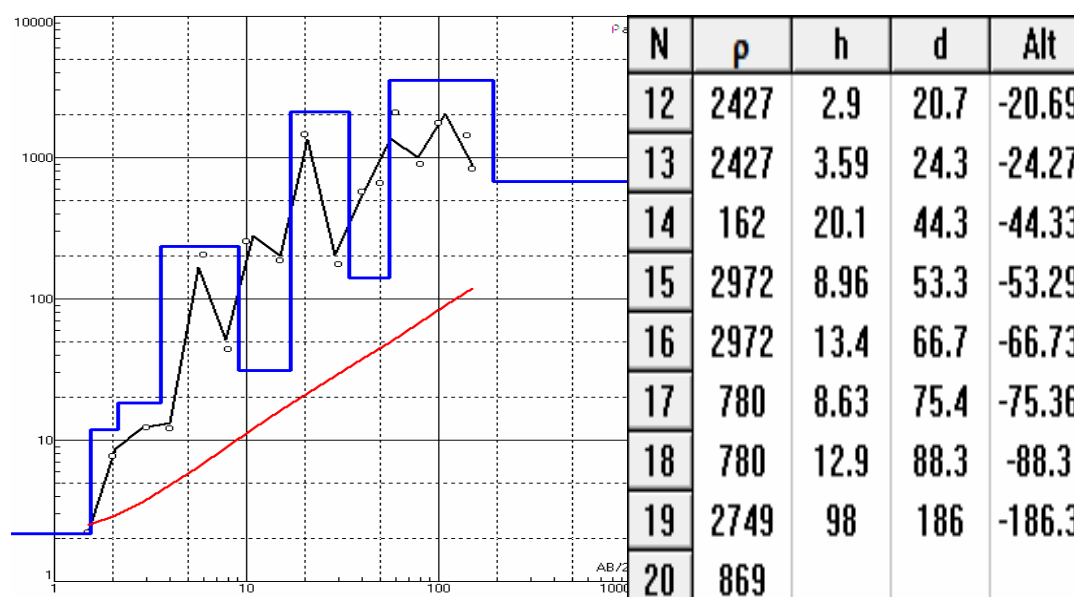
#### V. RESULTS

The measurement of resistance and their corresponding apparent resistivity value for VES 1, VES 10, and VES 20 at the Dumpsite and VES 30 at the control site are presented in figure 1, 2, 3, 4, and 5 respectively.

#### Interpretation of VES data for dumpsite

The apparent resistivity values obtained from the field measurements were plotted against half current electrode spacing on a log-log graph sheet. The resulting data were iterated to the lowest

root mean square (RMS) percentage error with the aid of the IPI2Win software (which uses raw data sounding interpretation method). An important step in the interpretation of resistive sounds survey data is to classify the apparent resistivity curves into types. In this research, the curves from the sites were predominantly type –H. This classification is made on the basis of the curves which depends on the number layers in the subsurface and the thickness of each layer. The uses of IPI2Win software which produce graphical representation of data grant the basis of making qualitative statement and observation of the study area. Figure 1, 2, 3, 4, and 5 shows the graph of apparent resistivity plot against half current electrode spacing for VES 1, VES 5, VES 10, VES 20, and VES 30 (Control) and the corresponding depth of each layer in the dumpsite.



**Fig 1: Graph of apparent resistivity plotted against half current electrode spacing for VES 1 dumpsite.**



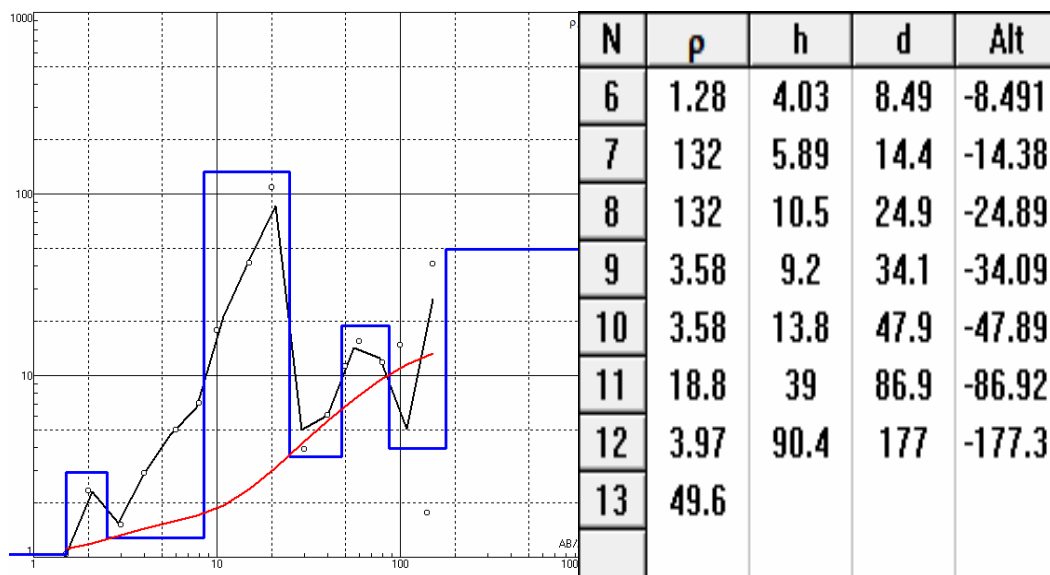


Fig 2: Graph of apparent resistivity plotted against half current electrode spacing for VES 5 dumpsite.

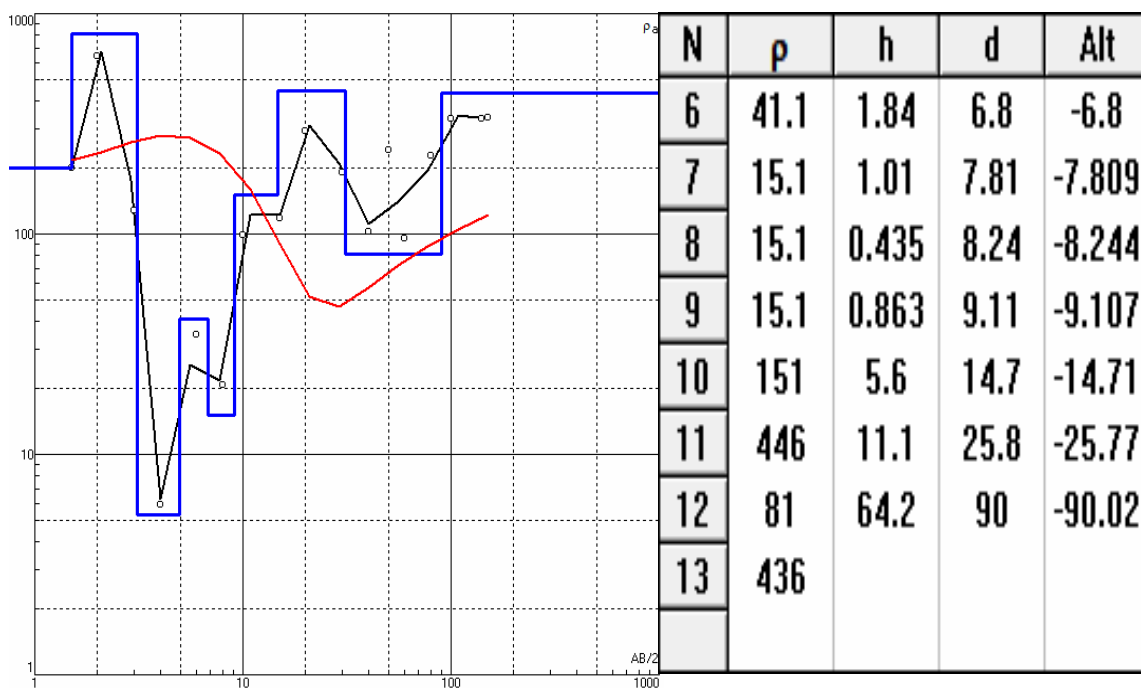


Fig 3: Graph of apparent resistivity plotted against half current electrode spacing for VES 10 dumpsite.

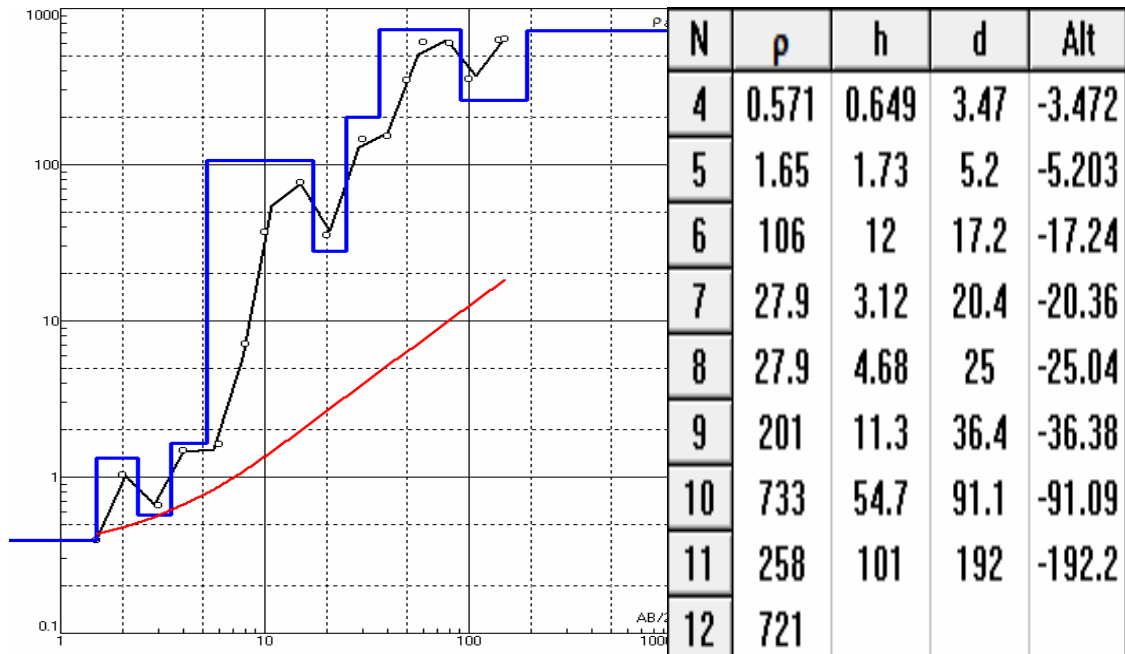


Fig 4: Graph of apparent resistivity plotted against half current electrode spacing for VES20 dumpsite.

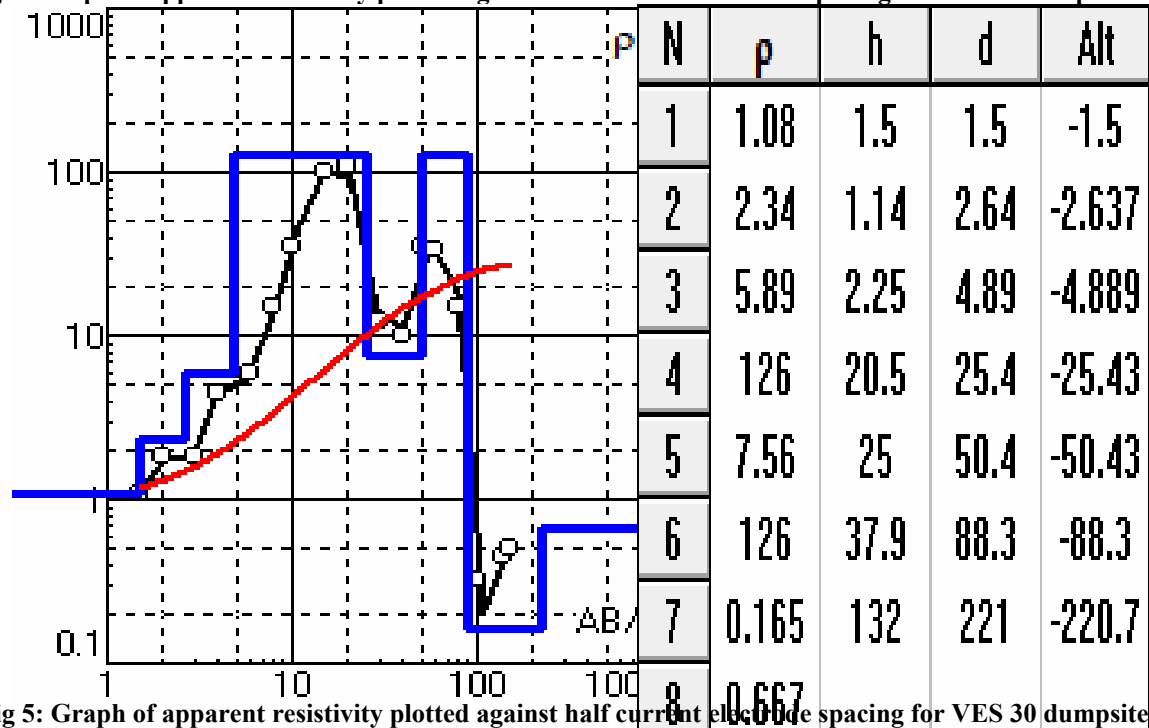


Fig 5: Graph of apparent resistivity plotted against half current electrode spacing for VES 30 dumpsite.  
Result of the 2-D Inversion for Profile line 1

Figure 6 shows the resistivity inversion results (iteration 3, 5.25% total average RMS error) for profile 1. Thus, indicating that, there is good fit between the measured and calculated apparent resistivity data as achieved. The Apparent resistivity

( $\Omega m$ ) is plotted against pseudo-depth (m). The Profile is located at the Southern end inside the dumpsite and it is 150m long, and runs in the West to East direction. Low resistivity zones ( $< 7 \Omega m$ ) were isolated near the surface with depth between 10.5m

to 15.4m which indicates contamination of the topsoil. In this profile the depth to the bedrock is shallow, about 26m and extends from  $x=26\text{m}$  to 140m along the profile. Colour variations in the basement rock are indication of contacts between different rocks which can be interpreted as fractures. The Orange colour indicates the weathered basement with resistivity value ( $<34\Omega\text{m}$ ).

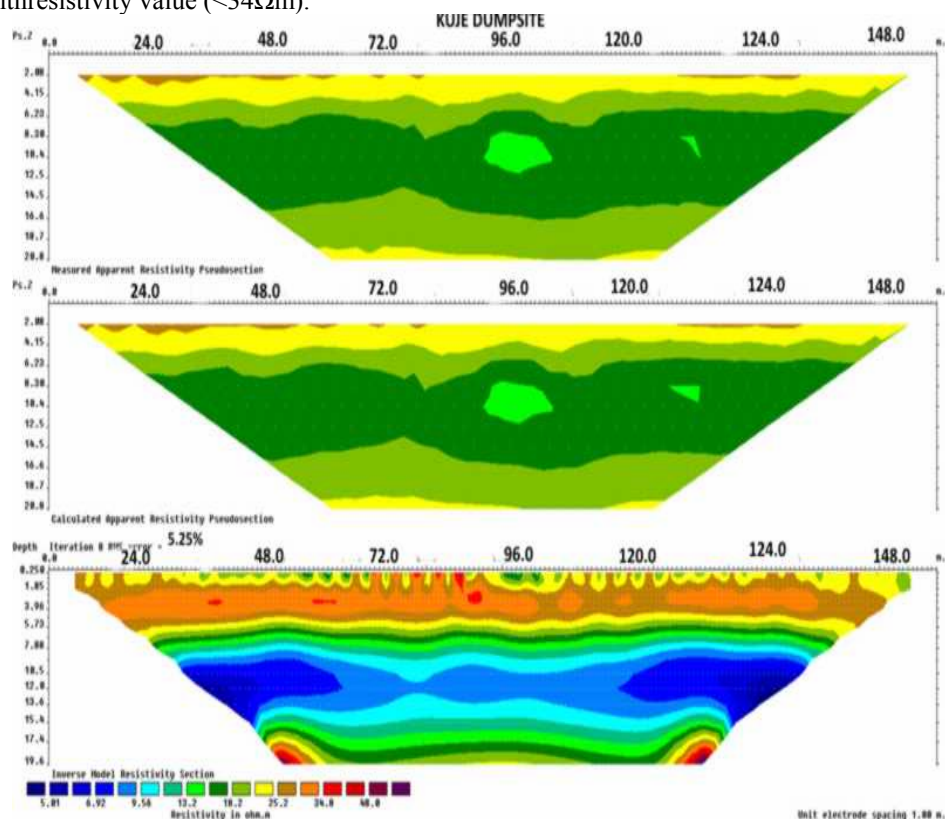


Figure 6: Result of 2D inversion of the Schlumberger-array data along profile line 1

### Result of the 2-D Inversion for Profile line 2

Figure 7 shows the resistivity inversion results (iteration 3, 1.72% total average RMS error) for profile line 2. Thus, indicating that, good fit between the measured and calculated apparent resistivity data were achieved. The Apparent resistivity ( $\Omega\text{m}$ ) is plotted against pseudo-depth (m). This Profile is located at about 100m away from profile 1 inside the dumpsite. This Profile runs in the West-South direction and it is 150m long. The materials in this profile are very resistive as shown by their resistivity values. The low resistivity ( $<27\Omega\text{m}$ ) zone, indicates the presence of leachate from the surface to a depth of 8m, which could be interpreted as topsoil contamination. The depth to the bedrock is also shallow, which is about

74m. The red and orange colour is the fresh basement as shown with resistivity values ( $>65\Omega\text{m}$ ).



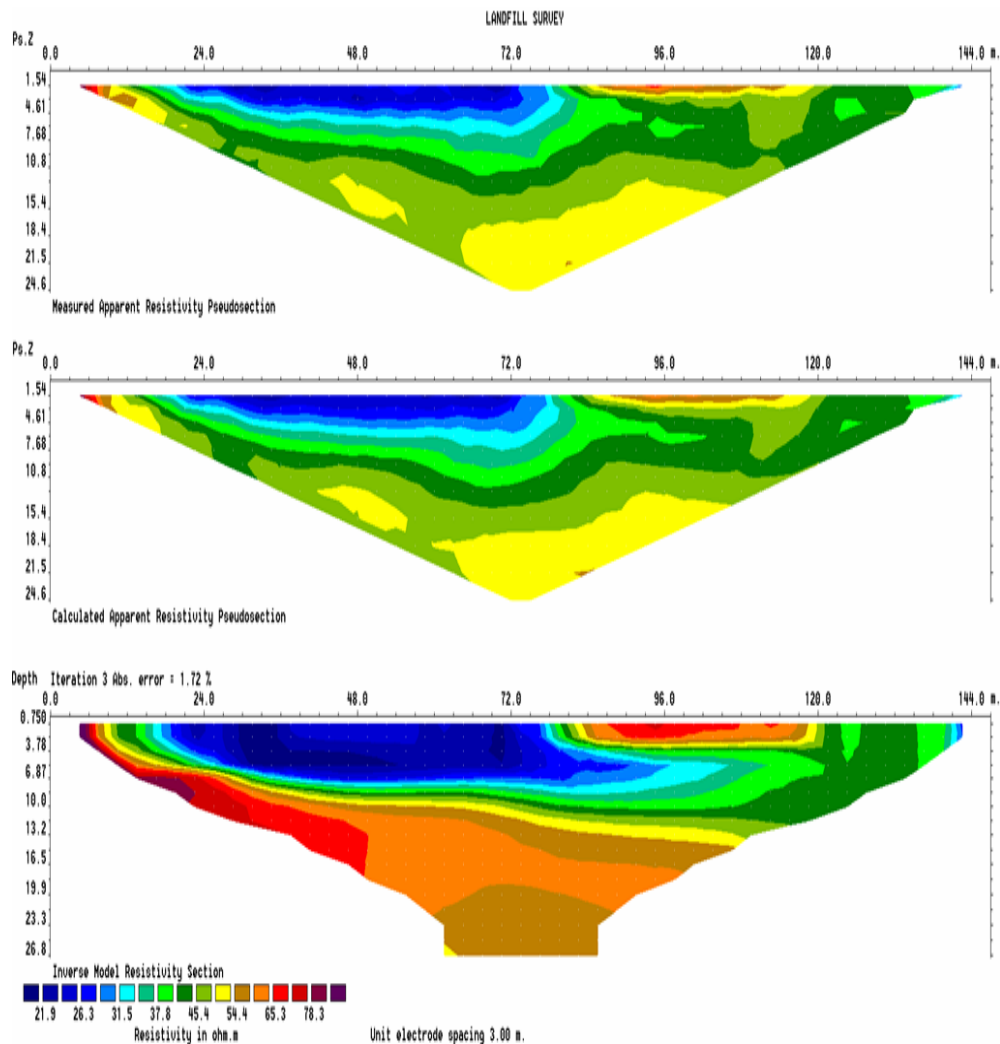


Figure 7: Result of 2D inversion of the Schlumberger-array data along profile line 2

### Result of the 2-D Inversion for Profile line 3

Figure 8 shows the resistivity inversion results (iteration 7, 5.1% total average RMS error) for profile line 3. Thus, indicating that, good fit between the measured and calculated apparent resistivity data were achieved. This profile is located at the Western end of the dumpsite and runs in the North-South direction of length 150m. Materials here, are less resistive as shown by the resistivity values ( $< 25 \Omega m$ ) of blue and purple colour. The low resistivity end ( $< 25 \Omega m$ ) could be attributed to contamination of the groundwater as a result of invasion of the leachate from  $x=50m$  to  $75m$  at  $1.84m$  depth and  $x=124.0m$  at  $2.38m$  depth. The migration of the leachate could be as a result of fractures (contacts between rocks of

different materials) or unconsolidated materials (sand or gravel).

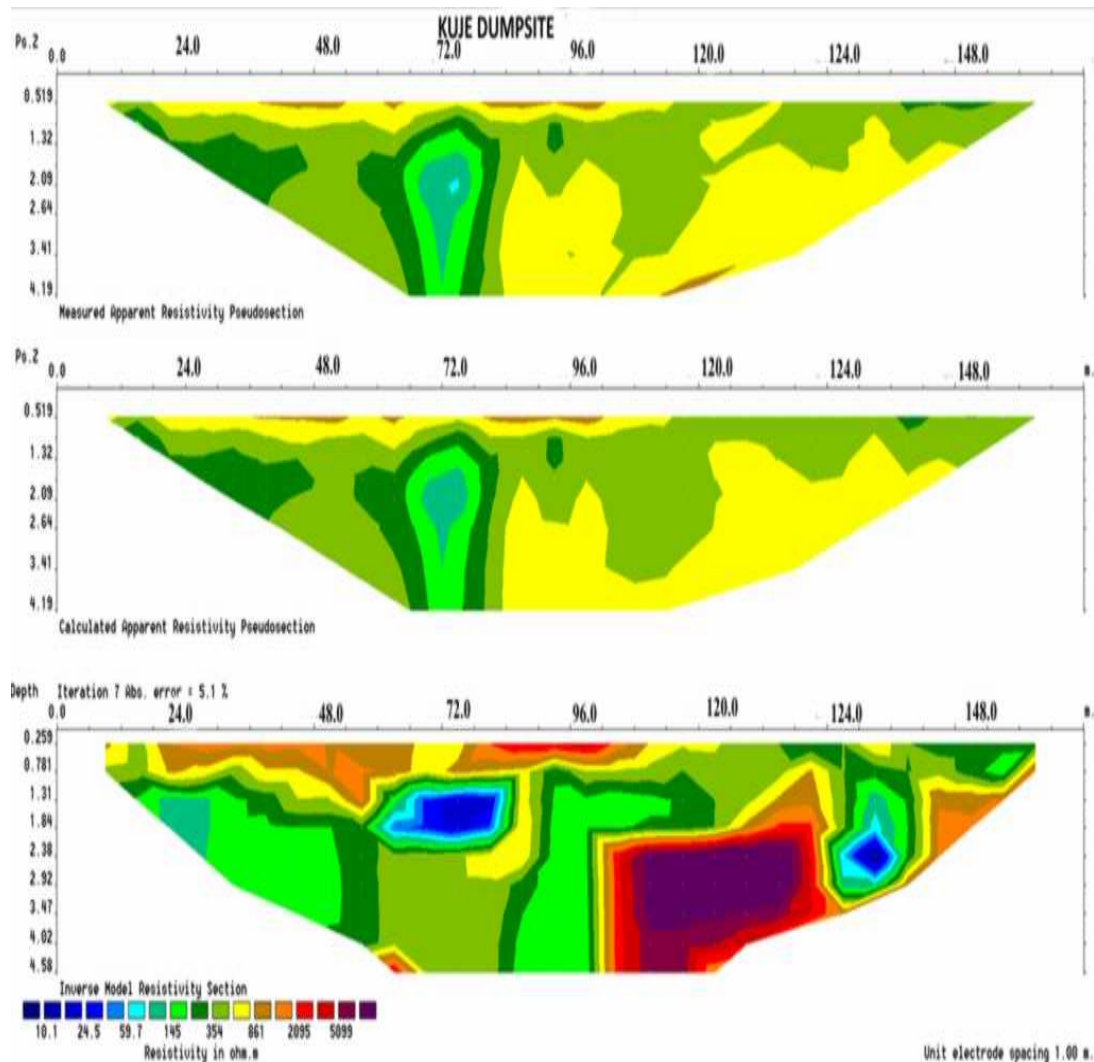


Figure8:Result of2D inversionoftheSchlumberger-arraydataalong profileline 3

#### Result of the 2-D Inversion for Profile line 4

Figure9showstheresistivity inversionresults(iteration3,6%totalaverage RMSerror)forprofile line 4. Thus, indicating that, good fit between the measured and calculated apparent resistivity data were achieved. Changing from deep blue to light blue also reflects the changes in the concentration of the leachate as it seeps down due to filtration by the sediments from  $x = 72\text{m}$  to  $148\text{m}$  and at a depth of  $24.9\text{m}$ . This Profile is located in the middle of the dumpsite. Low resistivity zones ( $< 6.98\Omega\text{m}$ ) are evident in this profile, indicating contamination of the topsoil and underground water. The light to deep blue colour along the profile indicates varying degree of concentration of the leachate. This pro

file displays materials that are unconsolidated (sand, gravel and fractured rocks). There is migration of the leachate which is believed to be due to fractures or unconsolidated subsurface material.

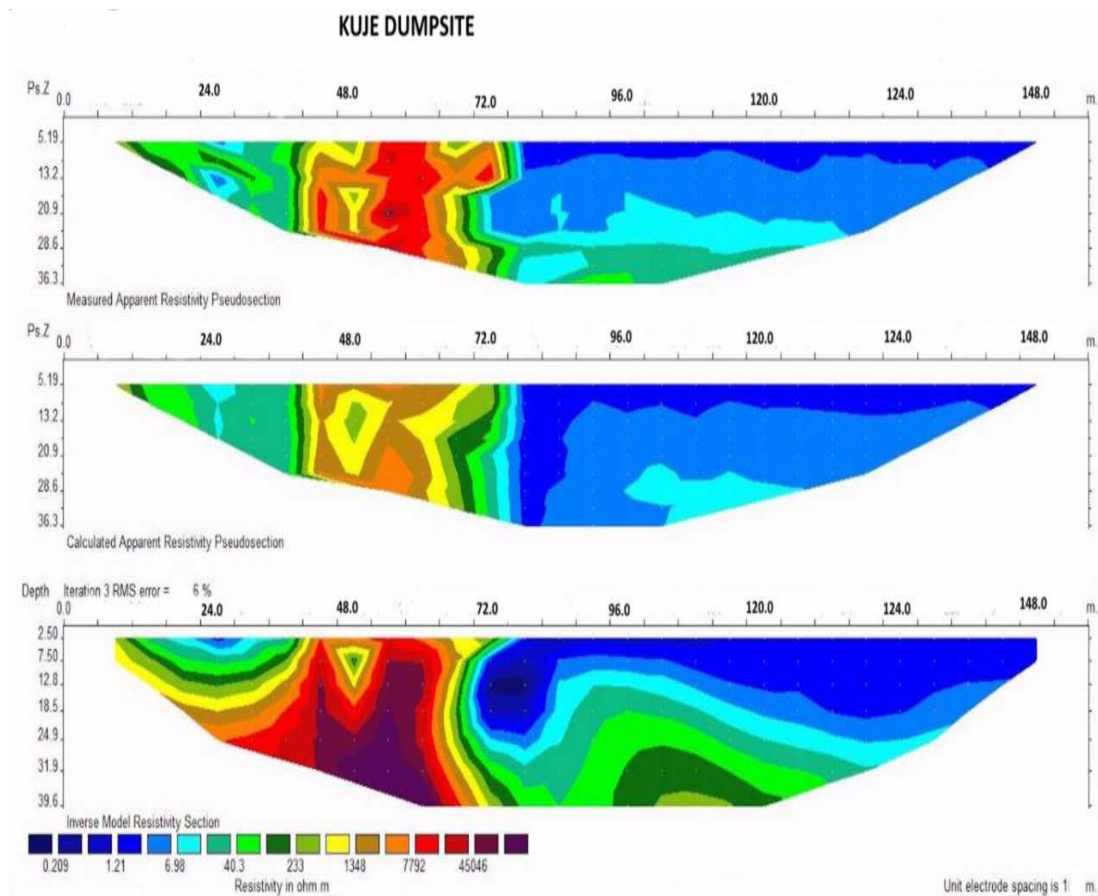


Figure9:Result of2D inversionoftheSchlumberger-arraydataalong profile line 4

### Result of the 2-D Inversion for Profile line5

Figure10showstheresistivity inversionresults(iteration3,5.1%totalaverage RMSError)forprofile line 5. Thus, indicating that, good fit between the measured and calculated apparent resistivity data were achieved. This Profile is located at 100m outside the dumpsite and runs in the North-South direction. This Profile is 100m long and is perpendicular to Profiles 1 and 2. There is evidence of contamination of the topsoil as shown by low resistivity value ( $<21.7\Omega m$ ) which occurs between  $x=50m$  to  $x=64m$  and from the ground surface to a depth of 2m. There is probably the presence of Clay which is impeding the downward movement of the leachates. This profile shows a partially weathered basement as indicated by the low resistivity value ( $<84.2\Omega m$ ).

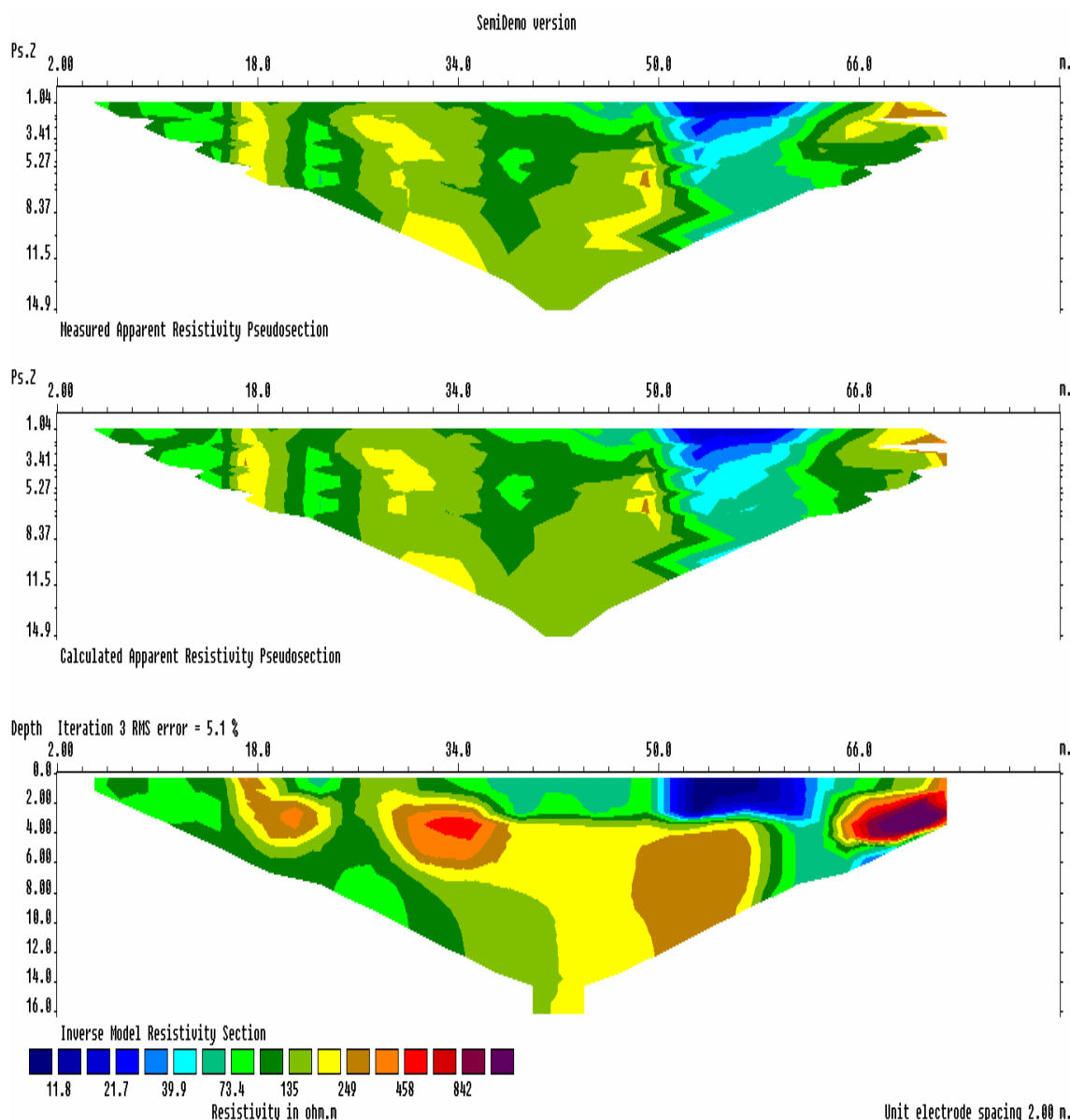


Figure 4.10: Result of 2D inversion of the Schlumberger-array data along profile

### 5 Physiochemical Analysis of Water Samples

Physiochemical analysis were carried out at Sheda Science and Technology Complex (SHETSCO) on water samples obtained from the well (SAMPLE A) and borehole (SAMPLE B) on and around the dumpsite for some selected parameters in order to ascertain the extent of contamination as presented in Table 2

Table 2 shows the physicochemical analysis result of the two major source of water in the study area which are hand dug well and bore hole. It shows the concentration of major ions present in the water samples and their percentage

compliance with World Health Organization (WHO) and Nigeria Standard for Drinking Water Quality (NSDWQ).

#### PHYSIOCHEMICAL ANALYSIS OF WATER SAMPLES

S/N	PARAMETERS	STANDARD WHO/NSDWQ	SAMPLE A	SAMPLE B
1	PH	7.0 – 8.0	6.90	6.89
2	TDS (mg/l)	500	1054	353.1
3	CONDUCTIVITY ( $\mu$ S/cm)	1000	2152	720.2
4	Fe <sup>2+</sup> (mg/l)	0.03	0.3253	0.1387
5	Zn <sup>2+</sup> (mg/l)	5	0.1352	0.0874
6	Mg <sup>2+</sup> (mg/l)	0.2	0.36	0.28
7	Cu <sup>2+</sup> (mg/l)	0.05	1.7	1.65
8	HARDNESS	100	771.68	1002.4

The PH value for both water samples lie within the permissible limit. The total dissolved solids (TDS) concentrations for hand dug well is 1054mg/l and 353.1mg/l for borehole. Hand dug water sample lies above the 500mg/l specified by WHO and NSDWQ while bore hole lies below the limit. Conductivity values of hand dug well is 2152 $\mu$ S/cm while borehole is 720.2 $\mu$ S/cm contrary to the standard limit of WHO and NSDWQ (1000 $\mu$ S/cm).

The result for total hardness for the hand dug well was given as 771.68 while that of bore hole was given as 1002.4. based on Sawyer and McCarthy (1967) classification for total hardness, 10% falls under 'moderate' class, 60% fall under 'hard' class while 30% fall under 'very hard'. For the two samples. Total hardness values were higher.

The two water samples were also tested for heavy metal such as iron, zinc, magnesium and copper. The result of the two water samples does not agree with WHO standard limit except for the value of iron in hand dug well water and that of magnesium in bore hole which approximately

agreed with the standard limit of WHO. The presence of high amount of those elements in the water samples are caused by the deposit of materials like metals irons etc on the dumpsite. Iron rusts and other materials decompose when come in contact with water when it rains. of the leachate which is believed to be due to fractures or unconsolidated subsurface material. In conclusion, the effect of leachate on groundwater repositories were assessed using vertical electrical sounding (VES).

The results of geoelectrical survey revealed that the study area is made up of three to four geoelectric layers and the fourth layer harbouring most of the aquiferous zones which is delineated as being more conductive and less resistive than the overlying layers as a result of seepage of leachate loaded contaminants into the layers. The results of the Physiochemical analysis revealed that almost all the analyzed parameters are not within the permissible limit of drinking water quality based on World Health Organization (WHO) standard and Nigeria Standard for Drinking Water Quality (NSDWQ) as shown in Table 4.2.



It could also be inferred that groundwater quality improves with increase in depth and distance from the dumpsite. The result from this study has thrown more light on the understanding of the impact of leachate on aquifer repositories around this dumpsite. It therefore suggests that soil analysis should be carried out to assess the effect of leachate on soil. Deeper drilling and constant monitoring of boreholes is also recommended. The Government should also consider the control of leachate generation, its treatment and subsequent recycling of wastes. This is necessary, since most aquifers in the study area are associated with leachate contamination. The continuous percolation of leachate into the subsurface constitutes health hazard and caution should be taken to prevent spread of waterborne diseases. Technical data and details on the state of the study area's subsurface have been made available by this research. These include the depth to groundwater and aquifer system, lithology, corrosivity of the topsoil, and the level of contamination at the dumpsite. It also showed the effects of a number of toxins coming from the dumpsites on the groundwater there as well as in the neighborhood.

We wish to express our deepest gratitude to the University of Abuja, and in particular, the Department of Physics, for their unwavering support and guidance throughout this project. Also like to extend our sincere appreciation to SHESTCO (Sheda Science and Technology Complex), where all laboratory tests were carried out. The exceptional facilities and support from the staff played a vital role in the success of this research.

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